



**HAL**  
open science

## Effect of seawater-freshwater cross-transplantations on viral dynamics and bacterial diversity and production

O. Bonilla-Findji, Emma Rochelle Newall, Markus G Weinbauer, M.D. Pizay, M.E. Kerros, Jean-Pierre Gattuso

► **To cite this version:**

O. Bonilla-Findji, Emma Rochelle Newall, Markus G Weinbauer, M.D. Pizay, M.E. Kerros, et al.. Effect of seawater-freshwater cross-transplantations on viral dynamics and bacterial diversity and production. *Aquatic Microbial Ecology*, 2009, 54 (1), pp.1-11. 10.3354/ame01256 . bioemco-00529292

**HAL Id: bioemco-00529292**

**<https://hal-bioemco.ccsd.cnrs.fr/bioemco-00529292>**

Submitted on 27 Oct 2010

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1  
2  
3 **Effect of seawater-freshwater cross-transplantations on viral dynamics and**  
4 **bacterial diversity and production**

5  
6 Osana Bonilla-Findji<sup>1</sup>, Emma Rochelle-Newall<sup>1,2\*</sup>, Markus G. Weinbauer<sup>1</sup>, Marie-Dominique  
7 Pizay<sup>1</sup>, Marie-Emmanuelle Kerros<sup>1</sup>, Jean-Pierre Gattuso<sup>1</sup>

8  
9  
10 **Address:** <sup>1</sup>: CNRS, Laboratoire d'Océanographie de Villefranche, B.P. 28, 06234 Villefranche-  
11 sur-mer Cedex, France ; Université Pierre et Marie Curie-Paris6, Laboratoire d'Océanographie  
12 de Villefranche, 06230 Villefranche-sur-Mer, France.

13 <sup>2</sup>: Present address: UR 103 Camelia, UMR 5119 ECOLAG, Université Montpellier II, Case 093,  
14 34095 Montpellier, France

15  
16  
17 \*For correspondence: email: emma.rochelle-newall@ird.fr

18 Tel : +33 (0)4 67 14 40 32

19 Fax. +33 (0)4 67 14 37 19

20  
21  
22  
23 **Keywords:** transplantation, bacteria production, bacteria diversity, virus  
24  
25  
26  
27  
28

29 **Abstract**

30

31 Dilution experiments were carried out to investigate the community composition and the  
32 metabolic response of seawater and freshwater bacteria to cross-transplantation, and the effects  
33 of non-indigenous bacterial hosts on viral dynamics. Changes in viral and bacterial abundance  
34 and production, as well as bacterial respiration, carbon demand and diversity were regularly  
35 monitored over a 6 day period. Bacterial production in the transplanted seawater and freshwater  
36 bacteria (SB-t and FB-t treatments) was stimulated up to 256% and 221%, respectively,  
37 compared to controls. The stimulation of bacterial production and carbon demand was  
38 accompanied by a decrease in bacterial richness. Net viral production was stimulated by 81% in  
39 SB-t and repressed by 75% in FB-t. Transplantation increased the virus-induced mortality of  
40 marine bacteria but decreased it for freshwater bacteria. These results suggest that (1) marine  
41 bacteria can readily oxidize freshwater dissolved organic matter (DOM), and (2) that freshwater  
42 viruses might be able to infect marine hosts, thus highlighting their potential role in fueling  
43 bacterial growth under resource stress or nutrient-depleted conditions.

## 44 **Introduction**

45           Substrate availability has been extensively studied as a potentially important factor  
46 controlling prokaryotic activity. It is well known that the quality and quantity of inorganic and  
47 organic nutrients can exert significant control on prokaryotic structure and function (e.g. Azam  
48 and Malfatti (2007)). In coastal areas, and particularly in estuarine systems, large variations in  
49 nutrient and organic carbon concentration can occur over relatively small distances. For  
50 example, in Randers Fjord, a small estuary (27 km long) in Northern Denmark, concentrations in  
51 inorganic and organic nitrogen increase by 1 to 3 orders of magnitude between the seawater and  
52 freshwater end-members, respectively (Veuger et al. 2004). Similarly, dissolved organic carbon  
53 concentrations also vary along estuarine gradients with generally higher concentrations in the  
54 freshwater end-members and lower concentrations in the seawater end-members (Fisher et al.  
55 1998, Abril et al. 2002). Superimposed upon these gradients, biological processes both modify  
56 and are modified by geochemical processes.

57           The dynamic nature of estuaries means that solutes and organic matter from the  
58 freshwater and seawater end-members become mixed along the estuarine gradient. The manner  
59 in which these inputs mix is related to the relative proportions of each input, as well as to other  
60 physical properties (tidal regime, winds, etc). Therefore, communities of bacteria and other  
61 planktonic organisms are exposed to strong physicochemical gradients and constantly varying  
62 environmental conditions along the length of the estuary. Moreover, in estuaries where  
63 freshwater inputs are relatively low compared to that of seawater inputs, such as in the Scheldt  
64 estuary, bacteria and other organisms experience a situation where salinity changes are reduced  
65 but where relatively large changes in dissolved organic matter (DOM) and other solutes can  
66 occur. This results in the exposure of seawater bacteria to freshwater DOM. The reverse side of  
67 this, of course, is that in these systems with high seawater to freshwater inputs, freshwater  
68 bacteria will be subject to the dual problems of changing salinity and DOM through mixing with

69 seawater. Estuaries are therefore attractive systems to address the question of substrate  
70 availability as a controlling factor of prokaryotic activity (Jones et al. 2007).

71         Strong compositional and metabolic changes have been shown to occur in bacterial  
72 communities between the freshwater and saltwater portions of estuaries (Bouvier & del Giorgio  
73 2002, Kirchman et al. 2004) and several authors have proposed the existence of a unique  
74 estuarine community that is found in the mixing zone (e.g. Crump et al. 2004). Indeed, the  
75 domination of a unique estuarine community *vs.* a mixing community in the middle estuary  
76 largely depends on the water residence time (Bouvier & del Giorgio 2002, Crump et al. 2004,  
77 Kan et al. 2006). However, regardless of water residence time, the instability of the environment  
78 is accompanied by changes in bacterial metabolism with bacterial cells more dedicated to  
79 physiological maintenance than cell production (del Giorgio & Bouvier 2002). Moreover, a  
80 recent meta-analysis of bacterial diversity data has shown that salinity plays an important role in  
81 structuring bacterial communities (Lozupone & Knight 2007) and it also seems to be a  
82 determining factor in carbon substrate utilisation in estuaries (Thottathil et al. 2008).

83         Although several studies have recently investigated bacterial responses to changes in the  
84 supply of organic matter in estuaries (Stepanauskas et al. 1999, Findlay et al. 2003, Langenheder  
85 et al. 2004), few have considered the impact of the associated variation of the coexisting viral  
86 communities, another key factor controlling bacterial dynamics. It is now well established that  
87 virioplankton exerts a “top-down” pressure and is a significant mortality agent for heterotrophic  
88 bacteria. Virioplankton also plays a significant role in shaping the composition and controlling  
89 the diversity of its hosts (Thingstad 2000, Wommack & Colwell 2000) and up to 50% of  
90 bacterial mortality can be due to viral activity (Fuhrman & Schwalbach 2003, Weinbauer 2004,  
91 Bouvier & del Giorgio 2007). Viral activity therefore has important effects on bacterial  
92 processes in aquatic systems. Over and above the direct effect of viruses on infected prokaryotic  
93 cells, viral lysis can enhance the activity of non-infected prokaryotic cells, which benefit from

94 the release of organic matter by cell lysis (Middelboe et al. 1996). Recent work also highlighted  
95 the complexity of the interactions between viral and bacterial activity and showed that viral lysis  
96 does not always result in increased bacterial growth efficiency (Bonilla-Findji et al. 2008). These  
97 authors reported that BR was stimulated (up to 113%) in the presence of active viruses whereas  
98 BP and BGE were reduced (up to 51%) and suggested that viruses enhance the role of bacteria as  
99 oxidizers of organic matter and as producers of CO<sub>2</sub>.

100         Similar to that observed for bacteria, spatio-temporal changes in the virioplankton  
101 composition and structure have been observed in estuaries (Wommack et al. 1999) and different  
102 viral infection rates are known to occur along salinity gradients (Almeida et al. 2001). Burst size,  
103 and thus viral production, tend to increase as a function of increasing DOM and nutrient  
104 concentration (Bettarel et al. 2004, Parada et al. 2006) . It is therefore probable that changes in  
105 DOM in terms of quality and quantity play a role in determining viral activity in estuarine  
106 systems.

107         The present study was carried out in the Scheldt Estuary (SW Netherlands and NE  
108 Belgium), one of the most eutrophic estuaries in Europe (Wollast 1988) in order to extend  
109 previous results on the functional response of a seawater bacterial community to freshwater  
110 dissolved organic matter (Rochelle-Newall et al. 2004). Two objectives were addressed: (1) to  
111 determine the structural and metabolic response of seawater and freshwater bacterial  
112 communities to cross-transplantation and (2) to investigate how viral dynamics change during  
113 transplantation relative to changes in DOM concentration and host diversity.

114

115

116

117

## 118 **Methods**

119        **Experimental setup.** In order to assess the response of bacteria and viruses to mixing  
120 along an estuarine salinity gradient, water samples were collected in April 2003 from the  
121 seawater and freshwater end-members of the Scheldt estuary (Fig. 1, Table 1). The response of  
122 seawater bacteria to freshwater DOM and viruses and, conversely, the response of freshwater  
123 bacteria to seawater DOM and viruses were investigated using dilution experiments. This  
124 transplant method was chosen to experimentally mimic the dynamic mixing of water masses and  
125 hence bacterial and viral communities and DOM along the estuarine salinity gradient of the  
126 Scheldt estuary.

127        Bacteria and viruses from the two end-members of the estuary were inoculated in filtered  
128 water from the other site in a 1:18 volume ratio (Fig. 2). At each station, 18 L of < 0.2 µm filtrate  
129 water were distributed into two 20 L, acid washed and Milli-Q rinsed polycarbonate carboys and  
130 then inoculated with 1 L of either unfiltered seawater or unfiltered freshwater in order to obtain  
131 four treatments: transplanted seawater bacteria (SB-t), seawater control (SB-c); transplanted  
132 freshwater bacteria (FB-t) and freshwater control (FB-c). The salinity of the freshwater filtrate  
133 was adjusted to 30 with an artificial seawater (ASW; Guillard 1975) mix to minimize any effect  
134 of the salinity on the seawater bacterial community.

135        The four treatments were incubated for 6 d in the dark and at *in situ* temperature (7.6°C)  
136 and sub-samples were collected for bacteria and viral counts, bacterial production and diversity  
137 at six time points (0h; 17h; 43h; 91h; 139h; 157h). Samples for determination of dissolved  
138 organic carbon (DOC) were taken at the beginning and end of the experiment.

139  
140        **Sample analyses.** Viral abundance was measured in duplicate, 1 ml samples fixed with  
141 glutaraldehyde (0.5% final concentration, EM-grade; Merck) for 30 min at 4°C in the dark, flash

142 frozen in liquid nitrogen and stored at  $-80^{\circ}\text{C}$  pending analysis by flow cytometry as described by  
143 Brussaard (2004).

144 Viral counts were performed with the CellQuest-Pro software (Becton Dickinson and  
145 Company) after staining with SYBR GREEN I (Invitrogen: S-7567) that was pre-diluted in 0.2  
146  $\mu\text{m}$  filtered, autoclaved Milli-Q (final dilution of  $5 \times 10^{-5}$  the commercial stock). Three viral  
147 groups were discriminated from scatter plots of side scatter (SSC) versus green fluorescence  
148 (FL1). These groups were labeled according to their increasing fluorescence signal: V1 (low),  
149 V2 (medium) and V3 (high). Previous comparisons (including the studied environment) of the  
150 total counts obtained by this method with viral abundances determined by microscopy were very  
151 similar and differences were generally less than 20% (O. Bonilla-Findji, unpublished). Viral  
152 production and decay was calculated from the net changes in viral abundance between each time  
153 point (Bratbak et al. 1990). The values obtained should therefore be considered as conservative  
154 estimations of viral production and decay.

155 Bacterial abundance was measured by a direct count method using epifluorescence  
156 microscopy and DAPI-stained samples (Porter & Feig 1980). Inspection of filters during  
157 enumeration did not reveal that flagellates were present in significant numbers. Bacterial  
158 production was estimated from  $^3\text{H}$ -leucine incorporation in accordance with the method of  
159 Kirchman (1992) and as previously described elsewhere (Rochelle-Newall et al. 2004). Bacterial  
160 respiration was calculated based on standard dark-bottle  $\text{O}_2$  consumption rates. Replicate BOD  
161 bottles were incubated in the dark following the JGOFS protocol (Knap et al. 1996) and at  
162 distinct time points, oxygen concentration was titrated using an automated Winkler titration  
163 technique with a potentiometric end-point detection (Anderson et al. 1992). The respiration rate  
164 at each time point  $T$  and its standard error were determined by regressing  $\text{O}_2$  concentration  
165 against time during the intervals of incubation  $T$  to  $T+I$ .

166 Due to the difference in the time scales of the measurement of bacterial production (1h) and

167 bacterial respiration (17 to 48 h), bacterial carbon demand (= bacterial production + bacterial  
168 respiration) was calculated using bacterial respiration, expressed in carbon units assuming a  
169 respiratory quotient of 1. Mean bacterial production was determined during the same time  
170 interval as used to measure the rate of respiration using  $((BP_{T1} + BP_{T2})/2)$ ,  $BP_{T1}$  and  $BP_{T2}$  are  
171 bacterial production at the start and end of the respiration measurement, respectively.

172 Samples (10ml) for determination of dissolved organic carbon were filtered through  
173 combusted (450°C, overnight) glass fiber filters (Whatman GF/F) and sealed in pre-combusted  
174 (450°C for 4h) glass ampoules after adding 12  $\mu$ l of 85% v/v phosphoric acid ( $H_3PO_4$ ). Samples  
175 were stored in the dark at 4°C pending analysis. DOC concentration was determined as  
176 previously described (Rochelle-Newall et al. 2007) by the high temperature combustion method  
177 using a Shimadzu TOC-5000 analyzer. Certified reference materials (Hansell Laboratory,  
178 University of Miami) were also used to assess the performance of the instrument on and between  
179 measurement days (Sharp 2002).

180 For extraction of DNA from prokaryotic cells 50 to 150 ml samples were recovered on a  
181 0.2  $\mu$ m pore-size polycarbonate filter (47 mm diameter; Whatman) and kept frozen at  $-80^\circ C$   
182 pending analysis. Nucleic acids were extracted from the filters and purified as described  
183 elsewhere (Winter et al. 2001, Winter et al. 2004b). In contrast to the phenol-chloroform  
184 extraction step from the original protocol, nucleic acids were extracted with 4.5 M NaCl and  
185 chloroform. This modified procedure avoids manipulation of a toxic chemical and yields  
186 fingerprints identical to those obtained by the original method (data not shown).

187 PCR conditions and chemicals were as described in Schäfer et al. (2001). Briefly, 1 to 4  $\mu$ l  
188 of the nucleic acid extracts were quantified on an agarose gel using a DNA mass ladder  
189 (EasyLadder I; Bioline #BIO-33045). When sufficient DNA was detected it was used in 50  $\mu$ l  
190 PCR reactions (1.5 mmol  $L^{-1}$   $MgCl_2$ , 0.25  $\mu$ mol  $L^{-1}$  of each primer and 2.5U *Taq* polymerase;

191 Sigma; #D 5930) together with a positive and a negative control. A fragment of the 16S rRNA  
192 gene was amplified using the bacterial primer pairs 341F-GC/ 907R (Schäfer et al. 2001).

193 When sufficient PCR products were obtained, denaturing gradient gel electrophoresis  
194 (DGGE) was carried out as described by Schäfer et al. (2001). PCR products (500 ng) were  
195 separated into bands by electrophoresis for 18 h at 100 V on acrylamide/bis-acrylamide (8%)  
196 gels prepared using a denaturing gradient from 30% to 70% (urea and formamide). DGGE gels  
197 were photographed with a gel documentation system GelDoc EQ (Bio-Rad) after 15 min staining  
198 with a 10X SYBR Gold solution (pre-diluted in 0.2 µm filtered, autoclaved Milli-Q; Molecular  
199 Probes: # S11494). Analysis of band patterns between lanes of the same gel was performed with  
200 the Quantity One Software (Bio-Rad). Apparent bacterial richness and band intensity (measured  
201 as peak area) is considered as the number of detectable bands on the DGGE gels.

202 The Statgraphics Centurion XV software package (Statpoint Inc, USA) was used to test  
203 the relationships between the treatments and respective controls. After verifying that  
204 assumptions were met (Shapiro-Wilks' test), t-tests were used to determine if the effect of  
205 transplantation was significant relative to the appropriate control. Significance is given at the  $p <$   
206 0.05 level.

207

## 208 **Results**

209 The *in situ* physico-chemical and biological characteristics of the two sampling stations  
210 differed considerably. The freshwater (FW) station exhibited higher bacterial abundance,  
211 production, and richness, viral abundance and dissolved organic carbon concentration than in the  
212 seawater (SW) station (Table 1).

213

214 **Bacterial dynamics.** The initial bacterial abundance in the incubations was 5.2 and 8.3 x  
215 10<sup>5</sup> ml<sup>-1</sup> in the seawater and the freshwater bacterial treatments, respectively (Fig. 3A). This was

216 in comparison to an *in situ* abundance of  $7.4 \times 10^6$  and  $3.1 \times 10^6 \text{ ml}^{-1}$  for the freshwater and  
217 seawater stations, respectively (Table 1). Cell numbers continuously increased during the  
218 experiment and reached values 2 to 3 times higher at the end of the incubation. This resulted in  
219 maximum abundances that were always less than the *in situ* values by a factor of 2-3 (Table 1,  
220 Fig. 3A). The only exception was in the seawater control treatment (SB-c), where bacterial  
221 abundance declined to initial values at the endpoint.

222 Bacterial production increased throughout the incubation although controls and  
223 transplanted treatments showed different patterns (Fig. 3B). While a lag period occurred in the  
224 transplanted treatments, it greatly increased in the controls during the first 36 h. This trend  
225 changed thereafter and bacterial production strongly increased and was up to 3-fold higher in the  
226 transplanted treatments than in the controls. Thus, although bacterial production was low in the  
227 transplantations in the short term (<36h), a two-fold stimulation was observed after 48 h in SB-t  
228 and FB-t. At the end of the experiment the stimulation of bacterial production was nearly 3-fold  
229 higher in SB-t than in FB-t.

230 During the first 91h of the incubation, bacterial respiration was significantly lower in the  
231 transplanted treatments, relative to their respective controls (t-test,  $p < 0.05$ ). During the second  
232 part of the incubation (> 91h), respiration generally decreased as a function of time in all  
233 treatments (Fig. 4A). Across the 7 day incubation period, respiration rates were statistically  
234 significantly lower in FB-t than in FB-c (t-test,  $p < 0.05$ ). This is in contrast to the seawater  
235 bacterial transplantation (SB-t) and control (SB-c) where no significant difference (t-test,  
236  $p > 0.05$ ) in respiration was observed over the 7-day incubation period.

237 Bacterial carbon demand, which is the sum of BR and BP was relatively stable  
238 throughout the control incubations (ranging from  $0.10$  to  $0.48 \mu\text{mol C L}^{-1} \text{ h}^{-1}$ ) and increased  
239 slightly at the end of the experiment (data not shown). In contrast, in the transplanted treatments,  
240 it was initially close to zero ( $0.01 \mu\text{mol C L}^{-1} \text{ h}^{-1}$ ) but greatly increased after 2 d (up to  $0.88 \mu\text{mol}$

241 C L<sup>-1</sup> h<sup>-1</sup>) until the end of the experiment (data not shown). The stimulation of bacterial carbon  
242 demand was stronger in SB-t (up to 239%) than in FB-t (up to 112%; Fig. 4B).

243 The relative amount of dissolved organic carbon utilized during the incubation differed  
244 between the treatments (data not shown). There was a trend of higher consumption in the  
245 freshwater and seawater controls (22 and 13% of the initial DOC concentration, respectively)  
246 than in the corresponding transplanted treatments (9 and 3.5%, respectively).

247 The apparent bacterial richness (i.e. the number of bands on DGGE gels) during the  
248 incubation decreased in all treatments and this effect was detectable after 43 h or less (Fig. 5).  
249 There were significant differences between the transplanted treatments and their respective  
250 controls (t-test, p< 0.05). Both transplanted treatments showed a continuous decrease in apparent  
251 richness whereas it stabilized and even slightly increased towards the end in both control  
252 incubations. The largest loss in apparent richness occurred in SB-t, where a decrease of 16 bands  
253 (41%) at the endpoint of the experiment was observed, compared to 13 in FB-t and only 6 in  
254 both controls.

255 Transplantation effects, whether negative or positive, were always larger for SB-t than for  
256 FB-t. The relative intensity of bands differed strongly during the confinement but varied also  
257 between treatments (Fig. 6). For example, band no. 1 decreased from around 10% of the total at  
258 the start of the experiment to less than 5% towards the end of the experiments in most  
259 treatments. The relative abundance of bands 21, 23 and 27 was typically less than 5% at the start  
260 of the experiments increased over the course of the experiment, with band 21 showing large  
261 increases (up to almost 30%). Interestingly, in the FB-c the intensity of these three bands  
262 remained comparatively stable throughout the experiment.

263

264 ***Viral dynamics.*** The initial viral abundance was higher in the freshwater-diluted  
265 treatments (FB-c and SB-t) than in the seawater treatments (SB-c and FB-t) accounting for 10 x

266  $10^7$  and  $1.5 \times 10^7$  particles  $\text{ml}^{-1}$ , respectively (Fig. 7A). This is in comparison to *in situ* values of  
267 11.1 and  $1.3 \times 10^7$  particles  $\text{ml}^{-1}$  for the freshwater and seawater station, respectively (Table 1).  
268 Contrasting viral dynamics were observed between the treatments diluted with freshwater filtrate  
269 (FB-c and SB-t), and the treatments diluted in seawater filtrate (SB-c and FB-t). The abundance  
270 of freshwater viruses in FB-c and SB-t, varied more over time than the abundance of seawater  
271 viruses in SB-c and FB-t, (Fig. 7A) and strongly decreased during the first day of incubation (by  
272 33% and 42% in FB-c and SB-t, respectively). At the end of the experiment, the viral abundance  
273 in FB-c and SB-t was slightly lower than at  $T_0$  (15%). In FB-t viral abundance at the endpoint  
274 was similar to  $T_0$ , while in SB-c it was 60 % higher than the initial values.

275         Although viral production exhibited similar patterns during the control incubations, it was  
276 43% to 235% higher in FB-c than in SB-c. Total viral production over the entire experiment was  
277 lower than decay in FB-c, FB-t and SB-t and viral production significantly exceeded the decay  
278 rate only in SB-c. By comparing the viral dynamics in FB-t and SB-t, relative to the seawater  
279 and freshwater controls (i.e. FB-t compared to SB-c and SB-t compared to FB-c), respectively, it  
280 is possible to evaluate the response of freshwater and marine viruses to the presence of a non-  
281 indigenous bacterial community (Fig. 7B). Despite initially high viral production values, viral  
282 production was repressed by up to 190% in the FB-t incubation, relative to the seawater control  
283 (SB-c). In contrast, the transplantation of a seawater bacterial community (SB-t) into freshwater  
284 viral community resulted in a strong stimulation of viral production (up to 840% after 91h).

285         The percentage of bacterial cells lysed per day can be estimated from the net increases in  
286 viral abundance in the incubations and assuming a burst size of 50 viruses, which represents high  
287 values from the North Sea (Winter et al. 2004a). Bacterial mortality due to viruses was higher in  
288 FB-c than in SB-c (23% and 12% cells lysed per day, respectively) and while transplantation of  
289 seawater bacteria increased virus-induced mortality to  $52\% \text{ d}^{-1}$ , the transplantation of freshwater  
290 bacteria decreased mortality to  $10\% \text{ d}^{-1}$ .

291           Most viruses (50-83%) were from the low fluorescence group (V1). However, the relative  
292 contribution of V1 to the total abundance differed between treatments (Fig. 7C). Over the course  
293 of the experiment, the contribution of the V1 group varied little in the SB-c and FB-t treatments.  
294 This is in contrast to the SB-t incubation, where the proportion of the V1 group was between that  
295 of FB-c (virus source) and SB-c (bacteria source) and exhibited an increasing trend towards the  
296 end of the incubation.

297 **DISCUSSION**

298 *Metabolic and structural response of bacterial communities to transplantation.*

299 This study investigates how the metabolism and diversity of estuarine and riverine  
300 bacteria responds to transplantation and to exposure to non-indigenous virus communities. The  
301 results show that transplantation increased both bacterial production and carbon demand in  
302 seawater as well as freshwater bacterial communities. This response is concomitant with a  
303 decrease in apparent bacterial richness. It suggests that bacteria were able to grow under  
304 allochthonous sources of dissolved organic matter supporting the conclusions of a previous study  
305 carried out in a Danish estuary (Rochelle-Newall et al. 2004). The data also show that  
306 transplantation stimulated total virus production and virus-induced mortality in SB-t but  
307 repressed it in FB-t.

308 It is well established that incubations can significantly affect bacterial community  
309 composition as well as activity parameters (Massana et al. 2001, Gattuso et al. 2002, Winter et  
310 al. 2004b). However, incubations are often the only possible approach to investigate ecological  
311 and biogeochemical issues. In the present study, the decrease in bacterial richness in the control  
312 incubations reflects both a confinement effect and the impact of the initial dilution. The decline  
313 of richness was similar in the two controls, where 6 bands were lost by the end of the  
314 experiment. In contrast, band loss was more than twice as high in the transplanted communities  
315 (13 and 16 bands for the FB-t and SB-t communities, respectively). This implies that the  
316 negative effects of transplantation on bacterial richness were higher than those of confinement  
317 alone. The lower apparent richness is probably related to the higher dominance of a few favored,  
318 fast growing phylotypes in the transplanted treatments relative to the controls. Furthermore, the  
319 loss of some of the phylotypes may well have helped to enhance the dominance of faster  
320 growing ones, such as can be observed in the SB-t incubation with loss of band 15 and the  
321 increase in band 23 intensity throughout the experiment.

322           These opportunistic species may have taken advantage of the nutrient amendment  
323 resulting from the dilution thus out-competing other members of the assemblage (Massana et al.  
324 2001, Gattuso et al. 2002, Winter et al. 2004b). Also, exposure of seawater bacteria to freshwater  
325 dissolved organic matter (FW-DOM) could have activated specific ectoenzymes or have favored  
326 certain members of the community capable of expressing them for hydrolyzing allochthonous  
327 DOM (Kirchman et al. 2004). This could explain the decrease in richness and the changes in  
328 community composition (Pinhassi et al. 1999).

329           The increase in bacterial production, respiration and carbon demand in the  
330 transplantations indicates that seawater bacteria could readily oxidize freshwater dissolved  
331 organic carbon, despite the fact that it has been reported as relatively less labile than marine  
332 DOC (del Giorgio & Davis 2003). This argument, combined with the relatively long residence  
333 time in the Scheldt estuary (1 to 3 months according to Wollast (1988), suggests that it is  
334 unlikely that there was an export of a labile FW DOC fraction that had not been completely  
335 taken up by the riverine bacteria as observed in other areas (Kirchman et al. 2004).

336           The higher metabolic activity and larger decrease in bacterial richness observed in SB-t  
337 compared to FB-t may also result from the different virus-induced bacterial mortality. Indeed,  
338 the virally-induced mortality was almost 5 times higher in the SB-t than in the FB-t. Moreover,  
339 although the change in salinity experienced by the FB-t treatment could have had a negative  
340 effect on the bacterial communities there was no salinity effect for the seawater bacteria  
341 transplanted into freshwater DOM as salinity was adjusted.

342  
343           ***Viral dynamics.*** The experimental setup (dilution of bacterial communities with  $< 0.2 \mu\text{m}$   
344 filtered water) artificially increased the initial virus to bacteria ratio. It was 5 to 9 times higher in  
345 the controls than in the corresponding inoculum. It increased by a factor of 9 in SB-t but was 5

346 times lower in FB-t as compared to FB-c. The FB-t bacteria therefore experienced reduced viral  
347 pressures relative to FB-c whereas SB-t bacteria experienced a stronger viral pressure than SB-c.

348 It has been shown that viral communities from different parts of an estuary exhibit  
349 pronounced differences in the genotypic composition (Wommack et al. 1999). Thus, the viral  
350 communities in the freshwater and seawater parts of the estuary should differ and this is  
351 potentially supported by the difference in the proportions of viral groups, as measured by  
352 fluorescence, between the two environments. There were differences between the FB-c and SB-c  
353 treatments potentially suggesting the presence of different viral communities in seawater and  
354 freshwater. These differences in relative proportion of low fluorescence viruses were also  
355 evident in the FB-t and SB-t incubations, reflecting the seawater and freshwater viral sources,  
356 respectively. The transplantation of seawater bacteria into a freshwater viral community resulted  
357 in large changes in the relative proportion of the low fluorescence virus group which may  
358 suggest that the seawater bacterial community produced viruses which differed from viruses  
359 produced by the indigenous freshwater bacterial community. However, this remains to be  
360 clarified in a more rigorous manner. Nevertheless, this hypothesis is also supported by the fact  
361 that the SB-t transplantation resulted in an elevated viral production rate (Fig. 7B). Thus, it is  
362 possible that the transplanted seawater bacteria were infected by viruses of the freshwater  
363 community.

364 Similar observations have been made by Sano et al. (2004). Interestingly, this pattern did  
365 not seem to hold for the seawater viruses in the FB-t incubation. Although viral production was  
366 high in the first hours, viral production was much lower in the transplanted incubation than in the  
367 control. This is suggestive of an increased survival of transplanted bacteria relative to that of the  
368 control and points towards the hypothesis that the freshwater bacteria did not act as hosts for the  
369 seawater viruses. Although the mechanisms for the lower viral production of the freshwater  
370 bacteria are not clear, the salinity effects of transplanting freshwater bacteria into virus-

371 containing seawater may play a non-negligible role in altering the virus-bacterial encounter and  
372 infection rates.

373 It should be noted that we compared treatments with similar initial abundances (FB-t with SB-c  
374 and SB-t with FB-c). Within these comparisons, total contact rate between viruses was 60%  
375 higher in the treatment with freshwater bacteria since initial bacterial abundance was higher in  
376 freshwater and the same % dilution was used for freshwater and seawater bacteria in the  
377 experiments. This could result in higher viral production rates by the freshwater bacteria  
378 (Murray & Jackson 1992). However, we found the opposite trend, i.e. transplantation of seawater  
379 bacteria into freshwater resulted in higher net viral production than for the freshwater bacteria,  
380 whereas transplantation of freshwater bacteria into seawater resulted in lower net viral  
381 production than the seawater control (Fig. 7B).

382         Viral production could be detected in the transplanted freshwater and the transplanted  
383 seawater bacterial community (Fig. 7A). Several non-mutually exclusive mechanisms could  
384 explain the production of viruses in transplanted bacterial communities. Firstly, the host range  
385 could be broader than previously assumed, allowing viruses to infect hosts from different  
386 environments. For example, Jensen et al. (1998) have argued that the concept derived from  
387 isolated virus-hosts systems that viruses do not trespass the genus barrier (Ackerman & Dubow  
388 1987) is an isolation artifact. This is also supported by Chiura (1997) who has shown that marine  
389 viruses can infect *Escherichia coli*. Secondly, transplantation into a different environment (e.g.  
390 with a large change in salinity) could cause induction of lysogenic cells (Jiang & Paul 1996).  
391 The enhanced growth rates (Fig. 3 and 4) could have acted as inducing agent (Weinbauer 2004)  
392 and caused the virus production observed. Finally, it is possible that cosmopolitan bacterial  
393 phylotypes that can grow and produce viruses in freshwater and marine conditions exist. Indeed,  
394 several identical bands were found in both environments and all treatments. Although there is no  
395 definitive evidence to support the hypothesis that freshwater viruses can infect seawater bacteria

396 and vice versa, it is clear that transplanted bacteria were able to produce viruses in the new  
397 environment, at least in the freshwater virus incubations. Moreover, as we found a strong  
398 increase in virus production, it is clear that these viruses came from either the transplanted  
399 community (by induction) or by new infection from the original viral community. Obviously,  
400 our estimates of viral production and decay can only be considered as net changes and so must  
401 be viewed as conservative estimates. Nevertheless, it is clear that these estimates were different  
402 between the different incubations (e.g. strong stimulation in SB-t and strong repression in FB-t)  
403 and so it is probable that there were real differences in viral production and decay between the  
404 different treatments. To conclude, although the freshwater viral community seems to be able to  
405 infect seawater hosts, the converse did not appear to occur.

406 **Acknowledgements**

407 We thank F. Gazeau for assistance, the captain and crew of the R. V. Belgica for their invaluable  
408 help and all the Eurotrophers for their camaraderie during the campaigns. This research was  
409 supported by the European Union in the framework of the EUROTROPH project (contract #  
410 EVK3-CT-2000-00040), the ATIPE grant to M.G.W. and a scholarship from the French  
411 Research Ministry to O.B.F. We also thank two anonymous reviewers whose comments greatly  
412 improved the manuscript.

413 **References**

- 414 Abril G, Nogueira M, Etcheber H, Cabecadas G, Lemaire E, Brogueira MJ (2002) Behaviour of  
415 organic carbon in nine contrasting European estuaries. *Estuarine Coastal and Shelf*  
416 *Science* 54:241-262
- 417 Ackerman HW, Dubow MS (eds) (1987) *Viruses of prokaryotes.*, Vol. CRC Press Boca Raton
- 418 Almeida MA, Cunha MA, Alcantara F (2001) Loss of estuarine bacteria by viral infection and  
419 predation in microcosm conditions. *Microbial Ecology* 42:562-571
- 420 Anderson LG, Haraldsson C, Roger L (1992) Gran linearization of potentiometric Winkler  
421 titration. *Marine Chemistry* 37:179-190
- 422 Azam F, Malfatti F (2007) Microbial structuring of marine ecosystems. *Nature Reviews*  
423 *Microbiology* 5:782-791
- 424 Bettarel Y, Sime-Ngando T, Amblard C, Dolan J (2004) Viral activity in two contrasting lake  
425 ecosystems. *Applied and Environmental Microbiology* 70:2941-2951
- 426 Bonilla-Findji O, Malits A, Lefèvre D, Rochelle-Newall E, Lemée R, Weinbauer MG, Gattuso J-  
427 P (2008) Viral effects on bacterial respiration, production and growth efficiency:  
428 consistent trends in the Southern Ocean and the Mediterranean Sea. *Deep Sea Research*  
429 *Part II: Topical Studies in Oceanography* 55:790-800
- 430 Bouvier T, del Giorgio PA (2002) Compositional changes in free-living bacterial communities  
431 along a salinity gradient in two temperate estuaries. *Limnology and Oceanography*  
432 47:453-470
- 433 Bouvier T, del Giorgio PA (2007) Key role of selective viral-induced mortality in determining  
434 marine bacterial community composition. *Environmental Microbiology* 9:287-297
- 435 Bratbak G, Heldal M, Norland S, Thingstad TF (1990) Viruses as partners in spring bloom  
436 microbial trophodynamics. *Applied and Environmental Microbiology* 56:1400-1405
- 437 Brussaard CPD (2004) Optimization of procedures for counting viruses by flow cytometry.  
438 *Applied and Environmental Microbiology* 70:1506-1513
- 439 Chiura HX (1997) Generalized gene transfer by virus-like particles from marine bacteria.  
440 *Aquatic Microbial Ecology* 13:75-83
- 441 Crump BC, Hopkinson CS, Sogin ML, Hobbie JE (2004) Microbial biogeography along an  
442 estuarine salinity gradient: Combined influences of bacterial growth and residence time.  
443 *Applied and Environmental Microbiology* 70:1494-1505

444 del Giorgio PA, Bouvier TC (2002) Linking the physiologic and phylogenetic successions in  
445 free-living bacterial communities along an estuarine salinity gradient. *Limnology and*  
446 *Oceanography* 47:471-486

447 del Giorgio PA, Davis J (2003) Patterns in dissolved organic matter lability and consumption  
448 across aquatic ecosystems. In: Findlay S, Sinsbaugh R (eds) *Aquatic Ecosystems:*  
449 *Interactivity of Dissolved Organic Matter*. Academic Press, p 399-424.

450 Findlay SEG, Sinsabaugh RL, Sobczak WV, Hoostal M (2003) Metabolic and structural  
451 response of hyporheic microbial communities to variations in supply of dissolved organic  
452 matter. *Limnology and Oceanography* 48:1608-1617

453 Fisher TR, Hagy JD, Rochelle-Newall E (1998) Dissolved and particulate organic carbon in  
454 Chesapeake Bay. *Estuaries* 21:215-229

455 Fuhrman JA, Schwalbach M (2003) Viral influence on aquatic bacterial communities. *Biological*  
456 *Bulletin* 204:192-195

457 Gattuso J-P, Peduzzi S, Pizay MD, Tonolla M (2002) Changes in freshwater bacterial  
458 community composition during measurements of microbial and community respiration.  
459 *Journal of Plankton Research* 24:1197-1206

460 Guillard RRL (1975) Culture of phytoplankton for feeding marine invertebrates. In: Smith WL,  
461 Chanley MH (eds) *Culture of marine invertebrate animals*. Plenum Publishing  
462 Corporation, New York, p 29-55

463 Jensen EC, Schrader HS, Rieland B, Thompson TL, Lee KW, Nickerson KW, Kokjohn TA  
464 (1998) Prevalence of broad-host-range lytic bacteriophages of *Sphaerotilus natans*,  
465 *Escherichia coli*, and *Pseudomonas aeruginosa*. *Applied and Environmental*  
466 *Microbiology* 64:575-580

467 Jiang SC, Paul JH (1996) Occurrence of lysogenic bacteria in marine microbial communities as  
468 determined by prophage induction. *Marine Ecology Progress Series* 142:27-38

469 Jones MK, Warner E, Oliver JD (2007) Survival and in situ gene expression of *Vibrio vulnificus*  
470 at varying salinities in estuarine environments. *Applied and Environmental Microbiology*  
471 74:182-187

472 Kan JJ, Crump BC, Wang K, Chen F (2006) Bacterioplankton community in Chesapeake Bay:  
473 Predictable or random assemblages. *Limnology and Oceanography* 51:2157-2169

474 Kirchman DL (1992) Incorporation of thymidine and leucine in the Subarctic Pacific: application  
475 to estimating bacterial production. *Marine Ecology Progress Series* 82:301-309

- 476 Kirchman DL, Dittel AI, Findlay SEG, Fischer D (2004) Changes in bacterial activity and  
477 community structure in response to dissolved organic matter in the Hudson River, New  
478 York. *Aquatic Microbial Ecology* 35:243-257
- 479 Knap AH, Michaels AE, Close A, Ducklow HW, Dickson AG (eds) (1996) Protocols for the  
480 Joint Global Ocean Flux Study (JGOFS) core measurements, Vol. Joint Global Ocean  
481 Flux Study, Scientific Committee on Oceanic Research, International Council of  
482 Scientific Unions, Intragovernmental Oceanographic Commission, Bergen, Norway
- 483 Langenheder S, Kisand V, Lindström ES, Wikner J, Tranvik LJ (2004) Growth dynamics within  
484 bacterial communities in riverine and estuarine batch cultures. *Aquatic Microbial  
485 Ecology* 37:137-148
- 486 Lozupone CA, Knight R (2007) Global patterns in bacterial diversity. *Proceedings of the  
487 National Academy of Sciences of the United States of America* 104:11436-11440
- 488 Massana R, Pedros-Alio C, Casamayor EO, Gasol JM (2001) Changes in marine  
489 bacterioplankton phylogenetic composition during incubations designed to measure  
490 biogeochemically significant parameters. *Limnology and Oceanography* 46:1181-1188
- 491 Middelboe M, Jorgensen NOG, Kroer N (1996) Effects of viruses on nutrient turnover and  
492 growth efficiency of noninfected marine bacterioplankton. *Applied and Environmental  
493 Microbiology* 62:1991-1997
- 494 Murray AG, Jackson GA (1992) Viral Dynamics - a Model of the Effects of Size, Shape, Motion  
495 and Abundance of Single-Celled Planktonic Organisms and Other Particles. *Marine  
496 Ecology-Progress Series* 89:103-116
- 497 Parada V, Herndl GJ, Weinbauer MG (2006) Viral burst size of heterotrophic prokaryotes in  
498 aquatic systems. *Journal of the Marine Biological Association of the United Kingdom*  
499 86:613-621
- 500 Pinhassi J, Azam F, Hemphälä J, Long R, Martinez J, Zweifel UL, Hagström Å (1999) Coupling  
501 between bacterioplankton species composition, population dynamics, and organic matter  
502 degradation. *Aquatic Microbial Ecology* 17:13-26
- 503 Porter KG, Feig YS (1980) The use of DAPI for identifying and counting aquatic microflora.  
504 *Limnology and Oceanography* 25:943-948
- 505 Rochelle-Newall EJ, Pizay MD, Middelburg JJ, Boschker HTS, Gattuso J-P (2004) Degradation  
506 of riverine dissolved organic matter by seawater bacteria. *Aquatic Microbial Ecology*  
507 37:9-22
- 508 Rochelle-Newall EJ, Winter C, Barron C, Borges AV, Duarte CM, Elliott M, Frankignoulle M,  
509 Gazeau F, Middelburg JJ, Pizay MD, Gattuso J-P (2007) Artificial neural network

510 analysis of factors controlling ecosystem metabolism in coastal systems. *Ecological*  
511 *Applications* 17:S185-S196

512 Sano E, Carlson S, Wegley L, Rohwer F (2004) Movement of viruses between biomes. *Applied*  
513 *and Environmental Microbiology* 70:5842-5846

514 Schäfer H, Bernard L, Courties C, Lebaron P, Servais P, Pukall R, Stackebrandt E, Troussellier  
515 M, Guindulain T, Vives-Rego J, Muyzer G (2001) Microbial community dynamics in  
516 Mediterranean nutrient-enriched seawater mesocosms: changes in the genetic diversity of  
517 bacterial populations. *FEMS Microbiology Reviews* 34:243-253

518 Sharp JH (2002) Analytical methods for total DOM pools. In: Hansell DA, Carlson CA (eds)  
519 *Biogeochemistry of marine dissolved organic matter*. Elsevier, New York, p 35-58

520 Stepanauskas R, Leonardson L, Tranvik LJ (1999) Bioavailability of wetland-derived DON to  
521 freshwater and marine bacterioplankton. *Limnology and Oceanography* 44:1477-1485

522 Thingstad TF (2000) Elements of a theory for the mechanisms controlling abundance, diversity,  
523 and biogeochemical role of lytic bacterial viruses in aquatic systems. *Limnology and*  
524 *Oceanography* 45:1320-1328

525 Thottathil SD, Balanchandran KK, Jayalakshmi KV, Gupta GVM, Nair S (2008) Tidal switch on  
526 metabolic activity: salinity induced responses on bacterioplankton metabolic capabilities  
527 in a tropical estuary. *Estuarine Coastal Shelf Science* 78:665-673

528 Veuger B, Middelburg JJ, Boschker HTS, Nieuwenhuize J, van Rijswijk P, Rochelle-Newall EJ,  
529 Navarro N (2004) Microbial uptake of dissolved organic and inorganic nitrogen in  
530 Randers Fjord. *Estuarine Coastal and Shelf Science* 61:507-515

531 Weinbauer MG (2004) Ecology of prokaryotic viruses. *FEMS Microbiology Reviews* 28:127-  
532 181

533 Winter C, Herndl GJ, Weinbauer MG (2004a) Diel cycles in viral infection of bacterioplankton  
534 in the North Sea. *Aquatic Microbial Ecology* 35:207-216

535 Winter C, Moeseneder MM, Herndl GJ (2001) Impact of UV radiation on bacterioplankton  
536 community composition. *Applied and Environmental Microbiology* 67:665-672

537 Winter C, Smit A, Herndl GJ, Weinbauer MG (2004b) Impact of virioplankton on archaeal and  
538 bacterial community richness as assessed in seawater batch cultures. *Applied and*  
539 *Environmental Microbiology* 70:804-813

540 Wollast R (1988) The Scheldt estuary. In: Salomons W, Bayne BL, Duursma EK, Förstner U  
541 (eds) *Pollution in the North Sea*. Springer Verlag, Berlin, p 183-193

542 Wommack KE, Colwell RR (2000) Virioplankton: Viruses in Aquatic Ecosystems.  
543 *Microbiology and Molecular Biology Reviews* 64:69-114

544 Wommack KE, Ravel J, Hill RT, Chun J, Colwell RR (1999) Population dynamics of  
545 Chesapeake Bay virioplankton: total-community analysis by pulsed-field gel  
546 electrophoresis. *Applied and Environmental Microbiology* 65:231-240

547

548

549

550

551

552

553

554

555

556  
 557 **Table 1.** Physico-chemical and biological characteristics of surface water sampled at the  
 558 freshwater (FW) and seawater (SW) ends of the Scheldt estuary. DOC: dissolved organic  
 559 carbon  $\pm$  standard deviation. BP: bacterial production  $\pm$  standard deviation. DGGE:  
 560 Denaturing Gradient Gel Electrophoresis.

561

	Stations	
	FW	SW
562 Temperature ( $^{\circ}$ C)	11.04	7.73
563 Salinity	1.20	30.4
564 DOC ( $\mu$ mol l $^{-1}$ )	427 $\pm$ 1.9	195.0 $\pm$ 1.6
Virus (x 10 $^7$ ml $^{-1}$ )	11.1	1.3
Bacteria (x 10 $^6$ ml $^{-1}$ )	7.4	3.1
BP ( $\mu$ mol C l $^{-1}$ h $^{-1}$ )	0.255 $\pm$ 0.006	0.070 $\pm$ 0.002
Bacterial richness (# of DGGE bands)	40	37

561

565 **Figure legends**

566

567 **Figure 1.** Map of the estuary and sample sites

568

569 **Figure 2.** Experimental set-up. Seawater or Freshwater (60 L) was filtered sequentially through  
570 GF/F (Whatman) filters and 0.2  $\mu$ m Durapore (Millipore) cartridge filters. 18 L of each filtrate  
571 was dispatched into four 20 L polycarbonate carboys. 1 L of unfiltered freshwater or seawater  
572 was added to each incubation to provide the natural bacterial inoculum. Two transplants and two  
573 controls were used: SB-c: seawater + seawater bacteria; SB-t: freshwater + seawater bacteria;  
574 FB-c: freshwater + freshwater bacteria; FB-t: seawater + freshwater bacteria.

575

576 **Figure 3.** Bacterial abundance (BA; panel A) and production (BP; panel B) in the different  
577 treatments: SB-c: seawater + seawater bacteria; SB-t: seawater bacteria transplanted in  
578 freshwater; FB-c: freshwater + freshwater bacteria; FB-t: freshwater bacteria transplanted in  
579 seawater. In panel B values represent mean  $\pm$  standard deviation

580

581 **Figure 4.** Bacterial respiration (BR) in the four treatments during the different phases of the  
582 experiment (panel A). Transplantation effect on bacterial carbon demand (BCD) across the  
583 incubation period in transplanted treatments (SB-t and FB-t) relative to their corresponding  
584 controls (panel B).

585

586 **Figure 5.** DGGE gels from the incubations. Each lane is labeled with the incubation and the time  
587 point. The bands discussed in the text are also noted. Upper image: time points T0 to T43, lower  
588 image T91 to T157. SB-c: seawater + seawater bacteria; SB-t: seawater bacteria transplanted in  
589 freshwater; FB-c: freshwater + freshwater bacteria; FB-t: freshwater bacteria transplanted in  
590 seawater. SD: standard.

591

592 **Figure 6.** Bacterial richness in the four treatments.

593

594 **Figure 7.** Viral abundance (VA; panel A), relative virus production (panel B) and relative % of  
595 V1 (low fluorescence group) in the different treatments across the experiment. Relative virus  
596 production is expressed as a % of the respective viral control (SB-c for the FB-t and FB-c for  
597 SB-t). The relative percentage of V1 is expressed as the % of the total abundance. SB-c:

598 seawater + seawater bacteria; SB-t: transplanted seawater bacteria; FB-c: freshwater +  
599 freshwater bacteria; FB-t: transplanted freshwater bacteria.

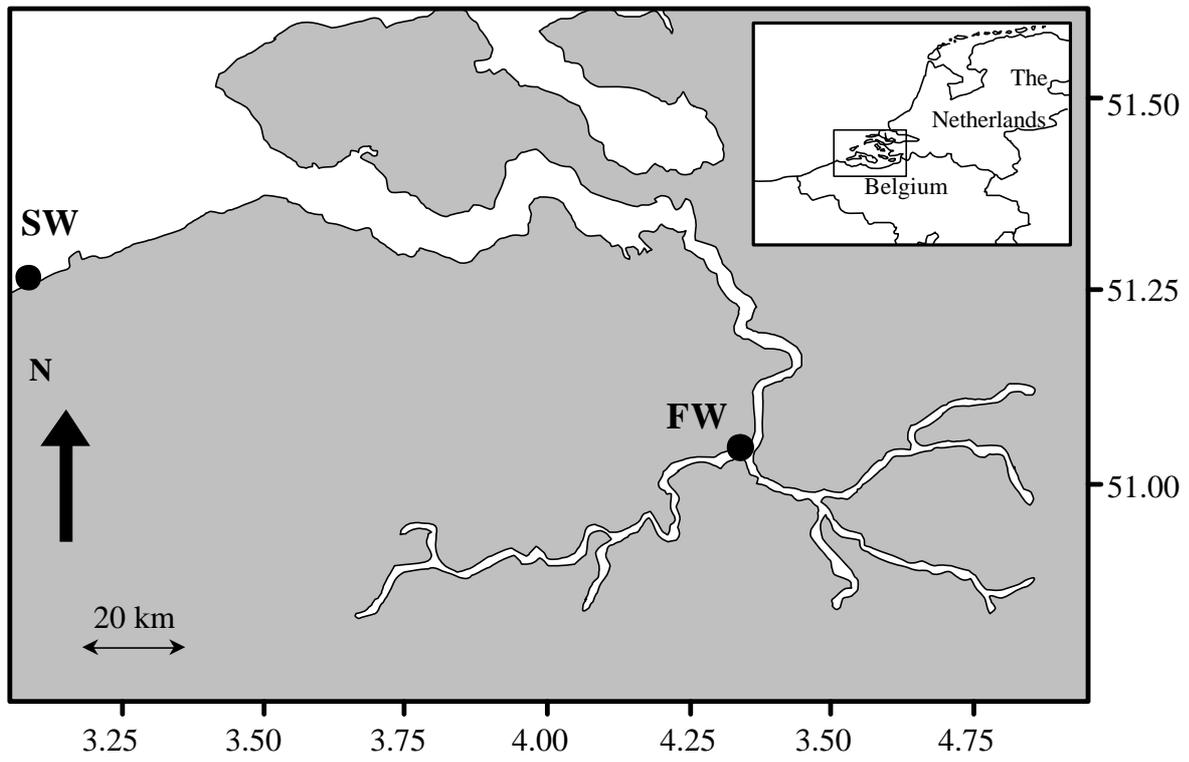


Fig. 1

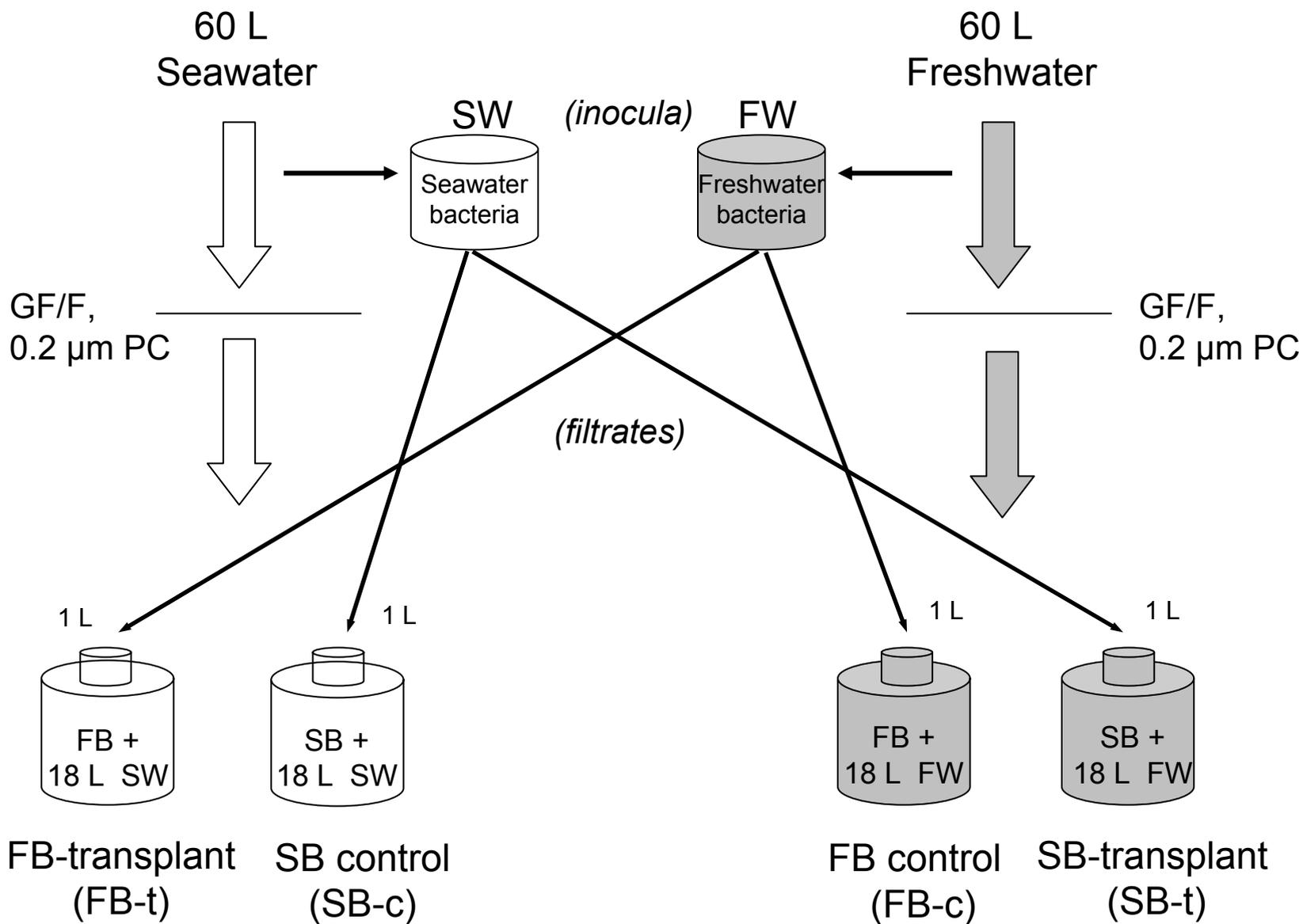


Fig. 2

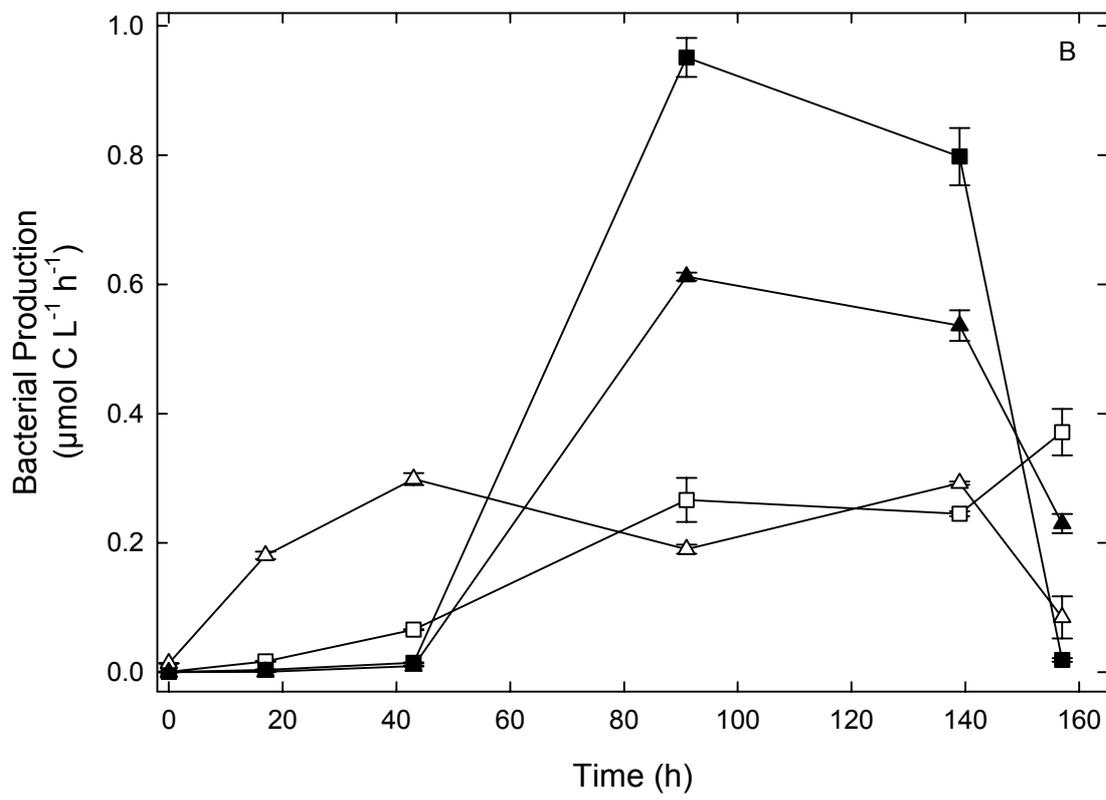
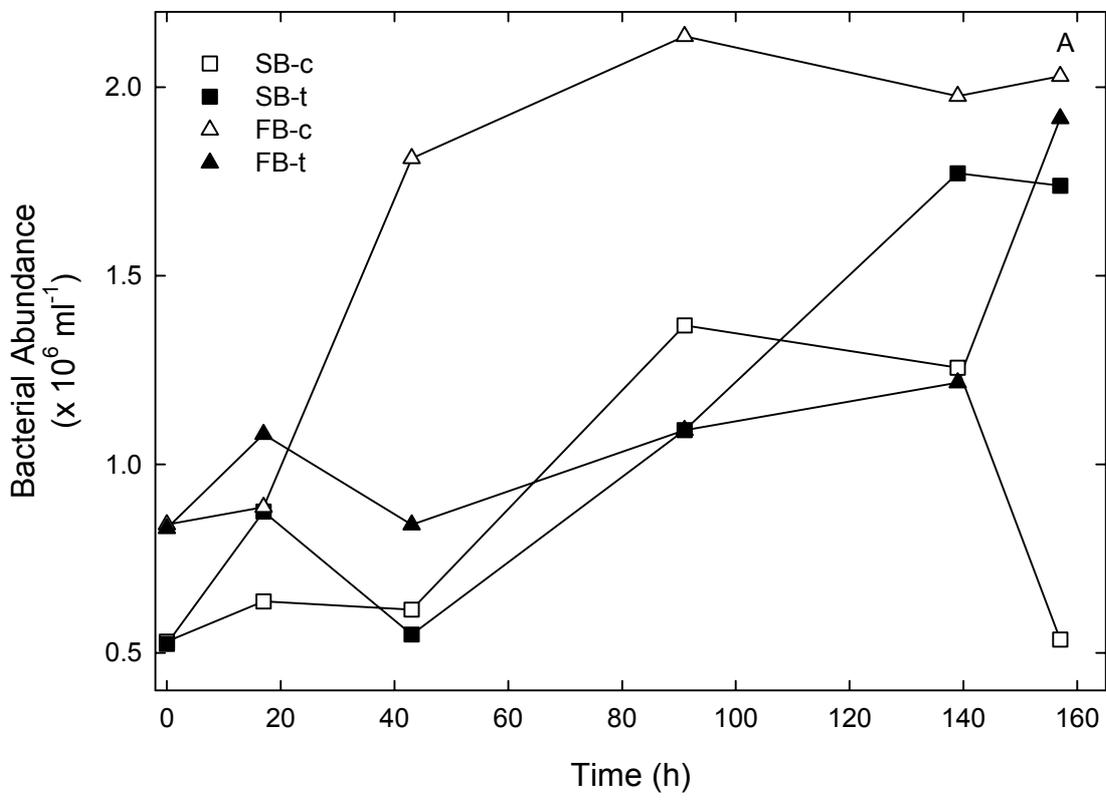


Fig. 3

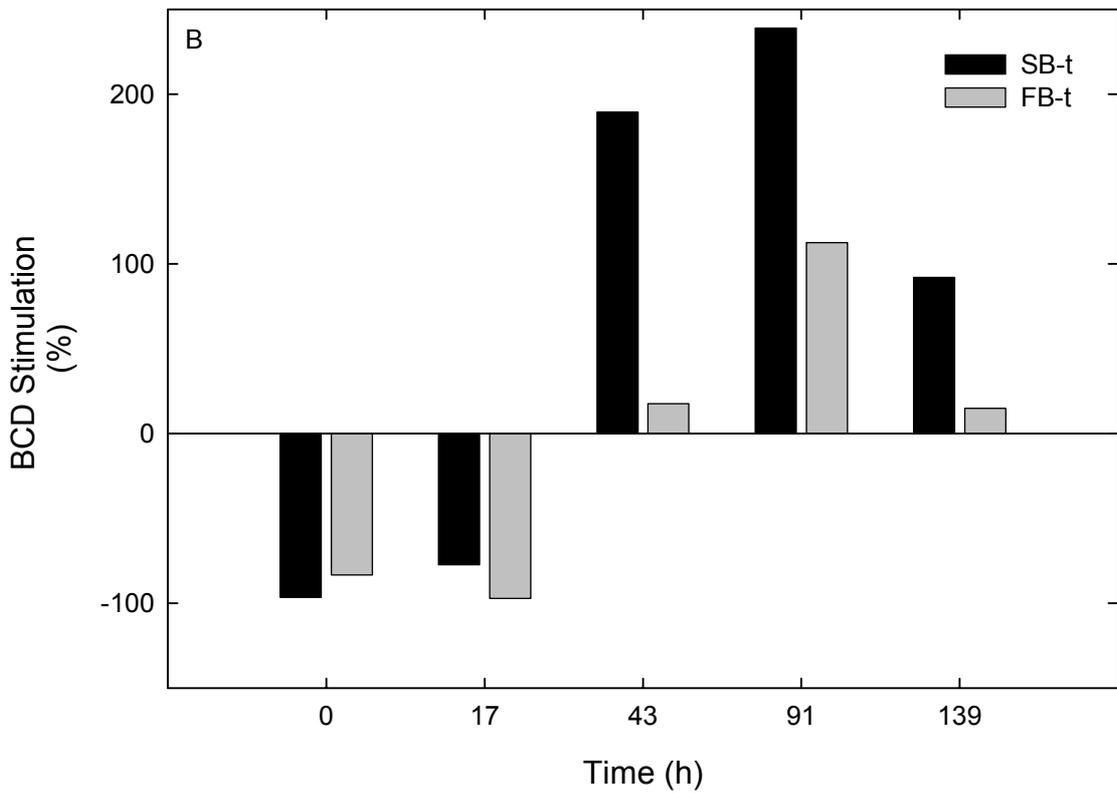
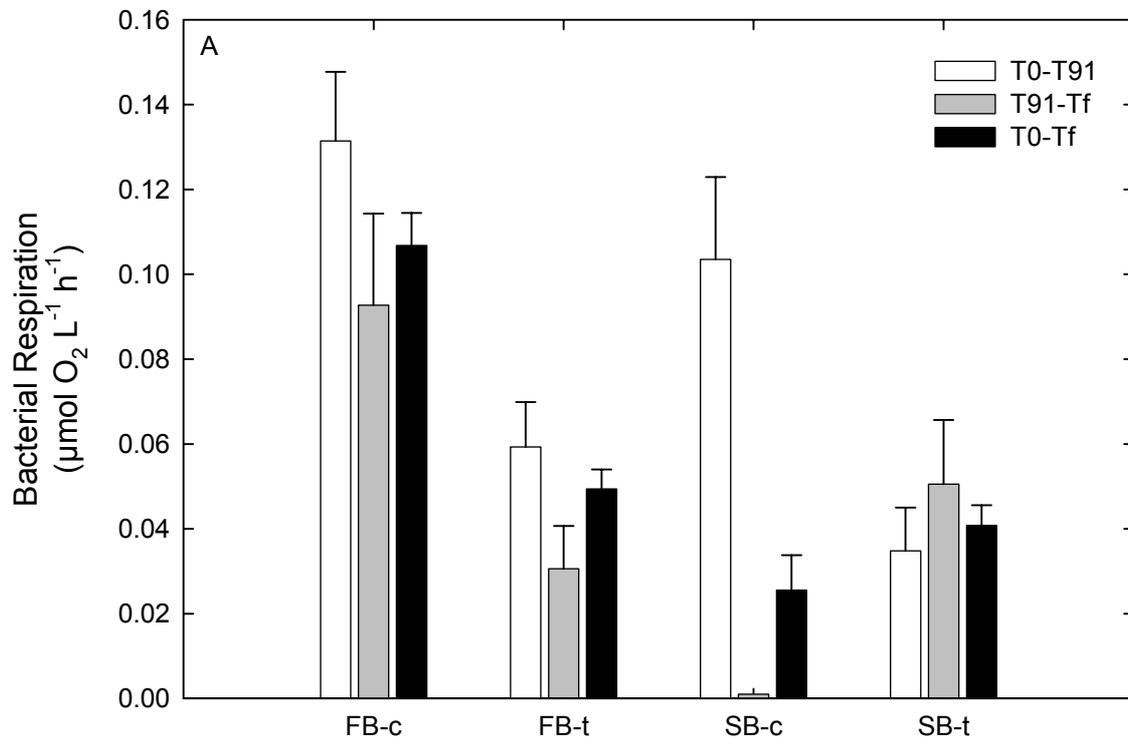


Fig. 4

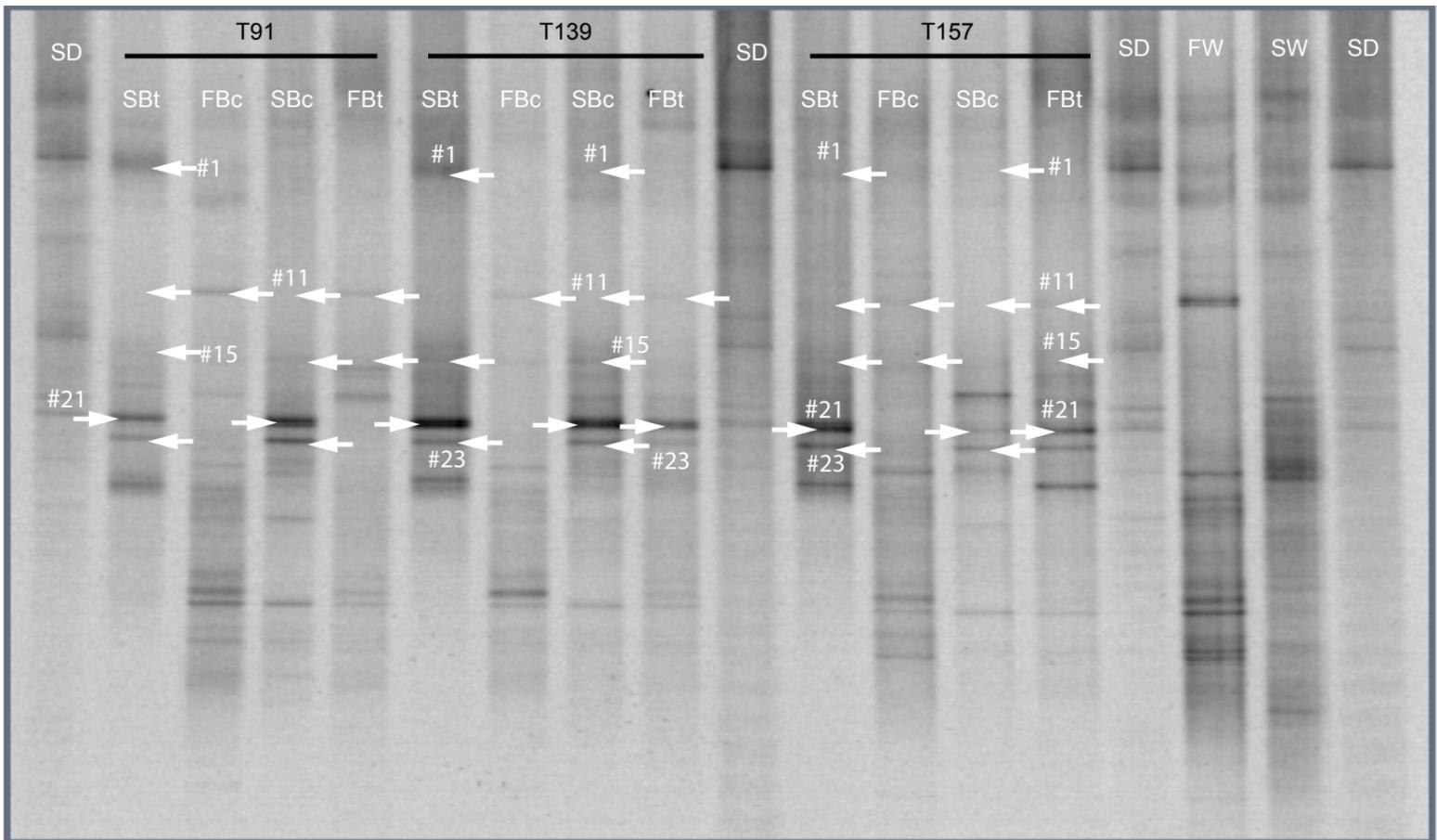
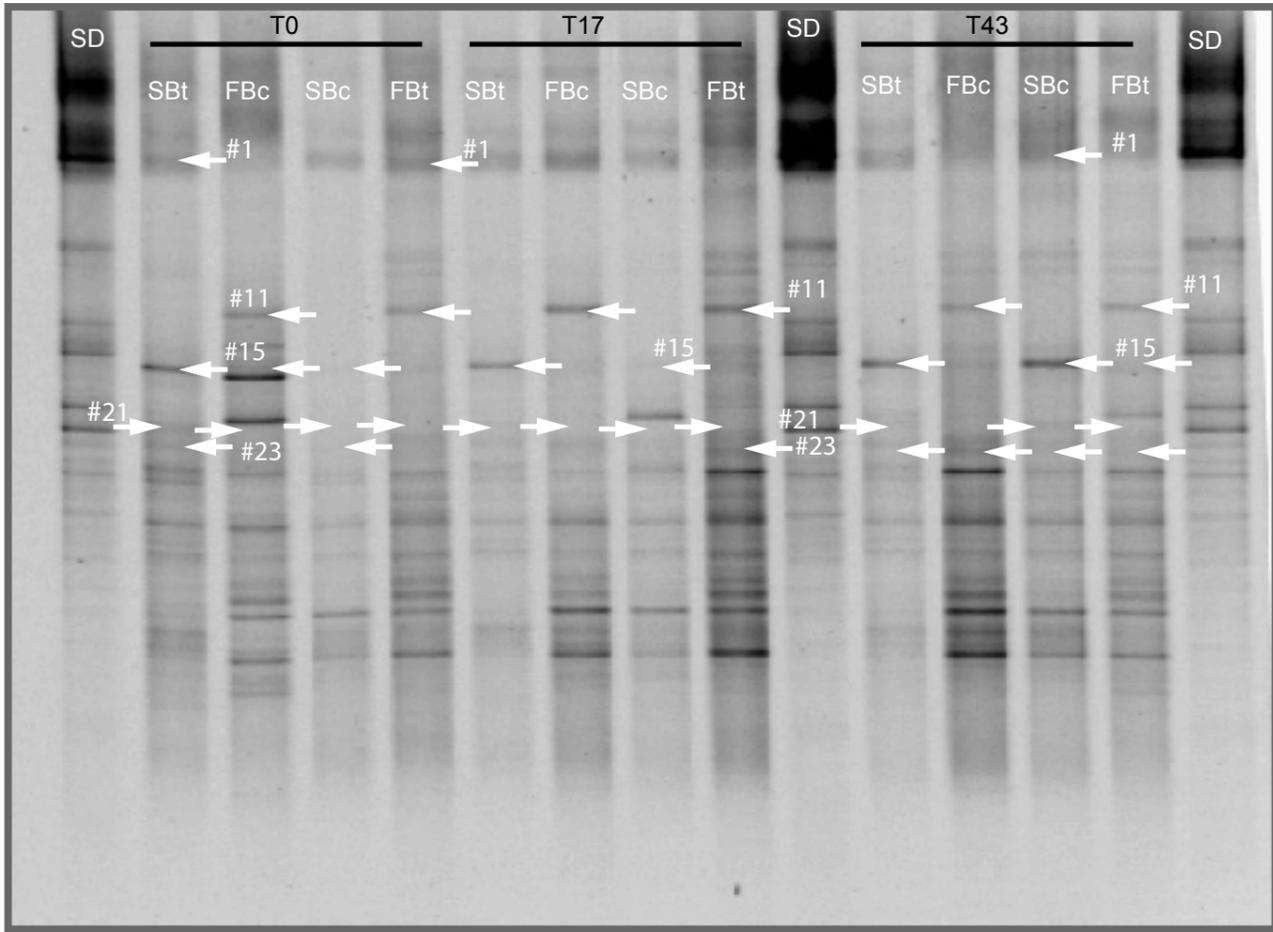


Fig. 5

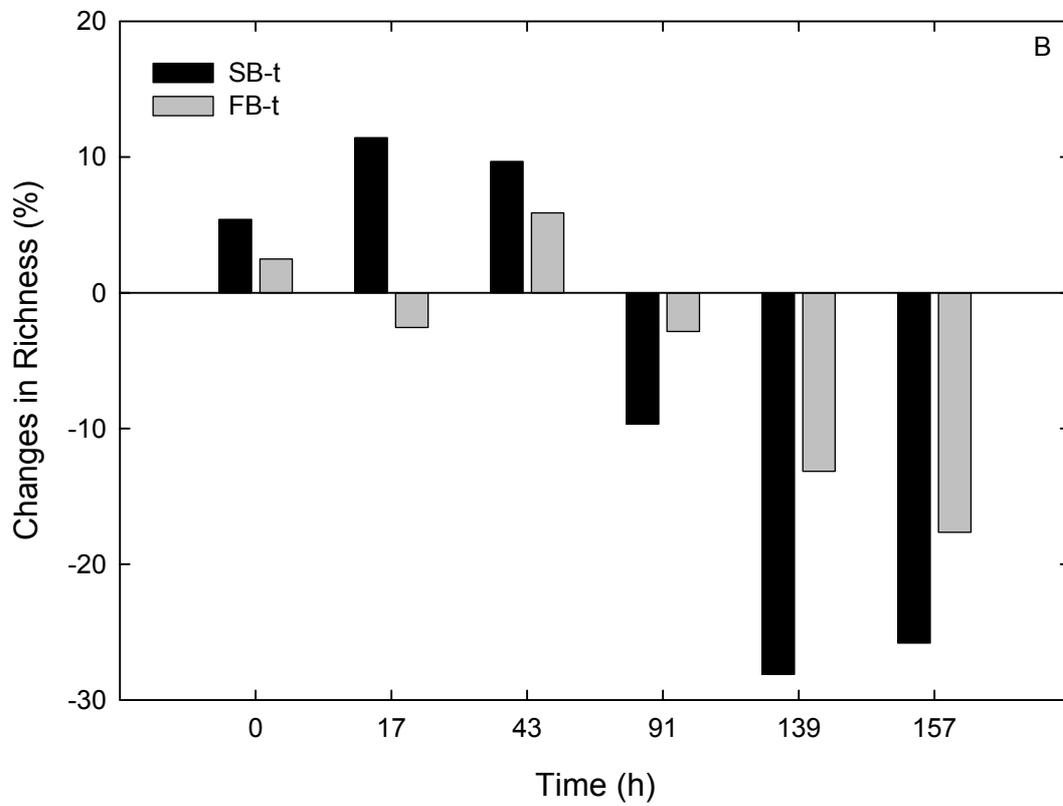
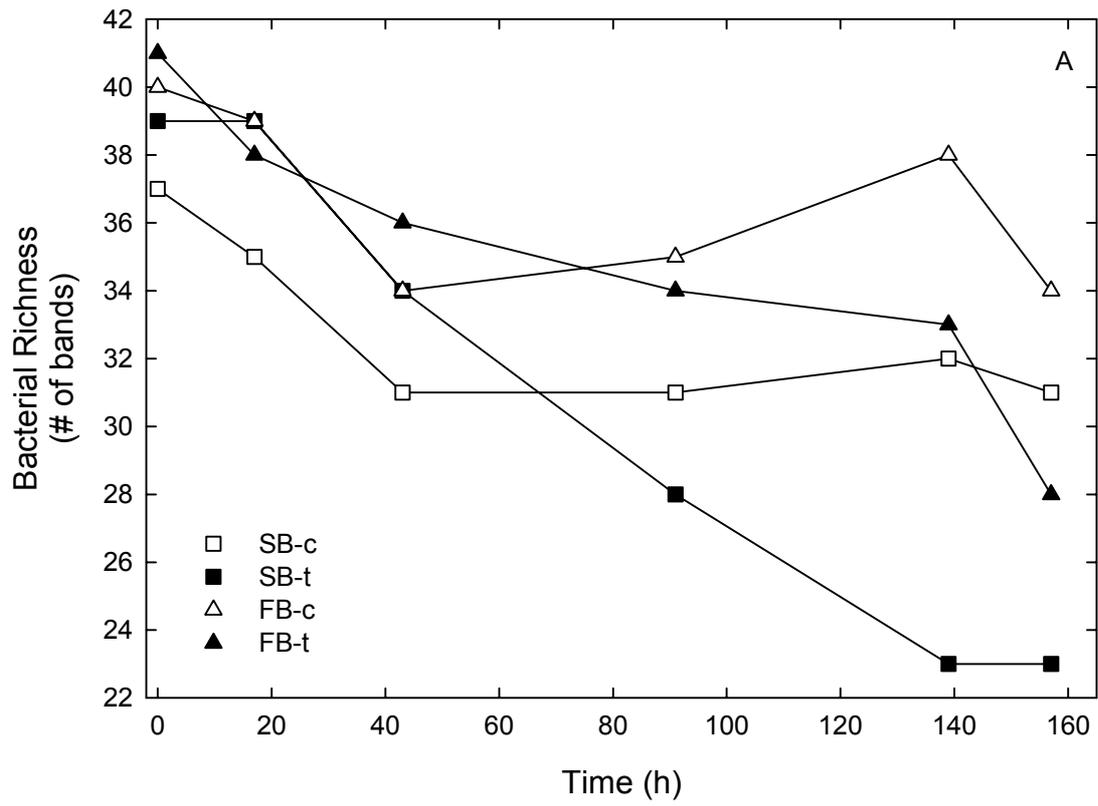


Fig. 6

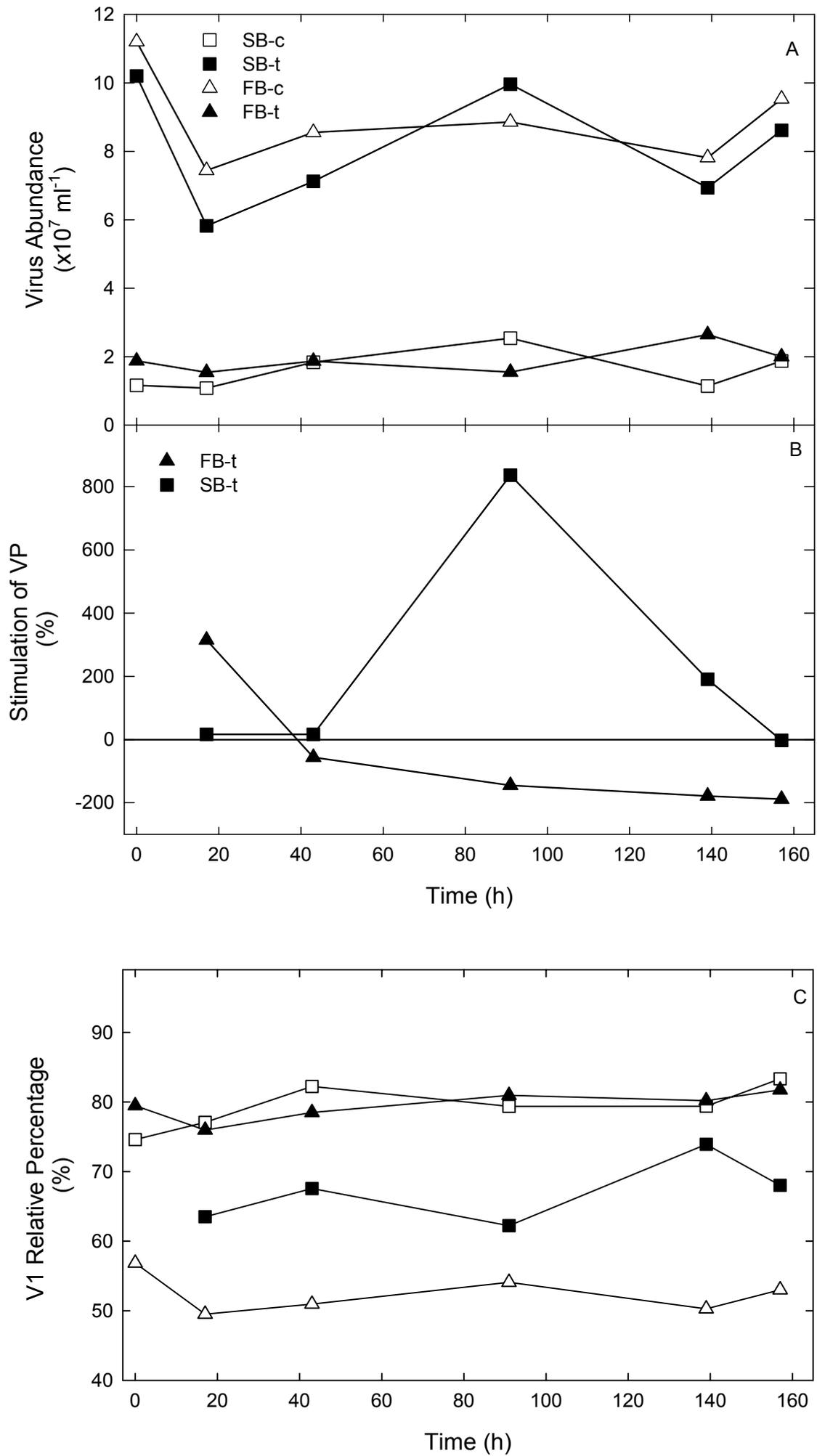


Fig. 7